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Radionuclide Concentrations in the Arkansas River Upstream and Downstream from the Nuclear I Power Generating Facility

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RADIONUCLIDE CONCENTRATIONS IN THE ARKANSAS RIVER UPSTREAM AND DOWNSTREAM FROM THE NUCLEAR I POWER GENERATING FACILITY

Technical Completion Report 1978

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INTRODUCTION

This report summarizes results obtained from a program designed to measure very low levels of some commonly produced radionuc1ides in the Dardene11e Lake area of the Arkansas River near the Arkansas Nuclear I Power Station operated by Arkansas Power and Light Company. The main thrust of this program was to determine the increase in the concentration of the radionuclides as a result of reactor operation as a function of their distance from the source. It was hoped to extend this study to include the effects of these emissions on the uptake of radionuc1ides into biological systems and their deposition in sediments.

Most studies of this kind are done by the isotopic gamma method of .. analyzing the gamma ray spectrum of a 0.5 - 3.5 liter sample of water usin a high-efficiency Ge(Li) detector. In many cases, especially in the areas of recently constructed reactors, the concentration of most radionuc1ides is below the detection limits for this system.

To increase the sensitivity of the determination of these nuclides, it was decided to perform radiochemical separations on a sample of approximately 20 liters in volume. Counting was done with a low-background, anti-coinci dence, gas-flow proportional counter. In the case of 58 Co and 137 Cs concentrations of greater than 0.5 pCi/1, low background NaI(T1) detectors and multichannel pulse height analysis were used for identification and confirmation.

The radionuclides whose concentrations were measured were 137 Cs, $144_{\text{Ce}}-144_{\text{Pr}}$, $90_{\text{Sr}}-90_{\text{Y}}$, 58_{Co} and 110_{mag} (see Table 1). After the Chinese nuclear test explosions in the autumn of 1976, 89 Sr and 141 Ce were added to the list. A gross beta activity measurement was also made for each sample.

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RELEASES

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The radioactive effluent from Arkansas Nuclear I consists of a low level, continuous discharge plus occasional planned releases of high activity waste that seemed to contain only 137 Cs and 58 Co in relatively large quantities along with small quantities of $^{110\rm{m}}$ Ag, 90 Sr- 90 Y, 134 Cs and 60 Co. Only in these planned releases were the latter two nuclides observed.

Samples taken after a planned release provide some information about short-term and long-term mixing in Lake Dardenelle.

EXPERIMENTAL PROCEDURES

A. Water Sampling

Samples varying in volume from $18 - 22$ liters were taken from the surface at four points each month (see Figure 1) except during periods of inclement weather. It was found that, in the outlet bay, water discharged from the cooling tower remained in a surface layer approximately two feet in depth. For all samples of 1976 and 1977, samples were taken monthly at stations 1, 2 and 4 and bimonthly, alternately, at stations 3 and 6. Sampling at station 6 was discontinued after March, 1977. $^{(1)}$

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B. Chemical Separations

- 1. An aliquot of 2.00 liters of river water was taken to analyze for $110m$ _{Ag}. An aliquot of 250 - 500 ml was taken to prepare the gross beta sample.
- 2. The remaining sample was acidified and carriers of $c\text{s}^+$, $c\text{o}^{2+}$, sr^{2+}
- (1) Alternate stations were used in November and December of 1975. Station 5 and 7 were used instead of 1, 21 instead of 3, and 16 instead of 2.

Figure 1: Locations of Sampling Stations

and Ce^{3+} were added. The sample was filtered through Whatman 42 paper after a settling period of two to seven days.

- 3. The sample was passed through a column of 100 g of Dowex-50 X8 at the rate of \sim 1 liter/hr.
- 4. The column was eluted with 500 ml of 6M HCl and 200 ml of water. The eluate was then evaporated to dryness.
- 5. The residue was dissolved in a minimum of conc. ${\tt HNO}_{\tt j}$ and the follo ing fractions were separated:
	- (a) Ce³⁺ as Ce(OH)₃ by the addition of conc. NH₃;

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- (b) \cot^{2+} as CoS by the addition of 1M thioacetamide;
- (c) Sr^2 as SrCO₃ by the addition of $(\text{NH}_4)^2$ ₂CO₃;
- (d) Cs⁺ remains in solution, which is evaporated and the residue is heated to remove NH_{λ}^{+} .
- 6. The Ce fraction, whose major impurity is $F e^{3+}$, was purified by the precipitation of Ce^{3+} as the oxalate, the oxidation of $Ce(III)$ to Ce(IV) followed by solvent extraction from the $9M$ HNO₃ phase into methyl isobutyl ketone. The Ce(IV) was reduced to Ce(III) and backextracted into water. The purified $G³⁺$ was precipitated as the oxalate and counted in that form (1).
- 7. The Co fraction, with numerous impurities, was purified by dissolving the sulfide in conc. HNO_3 , adding a Ni^{2+} holdback carrier and precipitating the mixed hydroxides with 10M KOH. After dissolving the hydroxides in glacial acetic acid, the $Co²⁺$ was precipitated as K_3 Co(NO₂)₆. The Co²⁺ is redissolved by the addition of conc. HCl and precipitated for counting as CoS (2).
- 8. The Sr fraction, whose major impurity is Ca^{2+} , was dissolved in 9M HNO₃

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and the Sr²⁺ was precipitated as $Sr(NO₃)₂$ by the addition of fuming HNO₃. The precipitate was dissolved in water and the Sr²⁺ was precipitated as the carbonate $(5,6)$. To assure that there was 2+ no Ca contamination, the precipitate was again dissolved in a minimum amount of $1\underline{M}$ HNO₃ and the Sr²⁺ was precipitated for counting as $SrS0_A$ by the addition of H_2SO_A .

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- 9. The Cs fraction was separated by coprecipitation with ammonium phosphomolybdate. After the ammonium, phosphate and molybdate ions were removed, the c_s^+ was precipitated as the perchlorate and counted in that form (3).
- 10. The 2.00 liter aliquot was filtered and A_g^+ carrier was added. Sufficient conc. NH_3 was added to redissolve the resulting chloride. After a short mixing period, conc. HCl was added to precipitate AgCl. The AgCl was dissolved in 1M NH₃ and the solution was scavenged twice with Fe(OH)₂. The solution was again acidified with conc. HCl and the resulting AgCl precipitate was filtered and counted.
- 11. The 250 ml aliquot was filtered and evaporated to dryness. The unfiltrable solids were slurried and quantitatively transferred to aluminum planchets. After they were dried without baking, they were stored in a dessicator until they were counted (4).

C. Counting Systems and Procedures

All the samples, except the gross beta samples, were wrapped in Mylar film with a thickness of $1-4$ mg/cm². These samples were counted in either a Tracerlab Omni-Guard system with a I" sample detector (G=0.209) or a Tennelec System equipped with a 2 1/4" Beckman anti-coincidence detector $(G=0.283)$.

The Cs and Co fractions from June and July, 1976 were also counted

with a $3''$ x $3''$ NaI(T1) detector housed in a 4" (minimum) lead shield. D. Analysis of Data

All beta counting data were corrected for systematic errors by using the expression:

$$
A_0(\text{pCi}) = \frac{R_0}{G \cdot f_c \cdot f_w \cdot f_b \cdot f_{ssa} \cdot 2.22} \tag{i}
$$

where A_0 = the activity (pCi) on the day of collection;

 R_0 = the count rate (cpm) on the day of collection;

 $G =$ the geometry of the detector;

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- f_c = the correction factor for absorption of radiation by the sample cover;
- $f =$ the correction factor for absorption of radiation by the w detector window;

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f_b
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 = correction factor for backscattering;
\n f_{ssa} = correction factor for sample self-absorption and scatter:
\n2.22 = dpm/pCi

The geometry factor was measured by determining the count rate of standardized samples of H , $\frac{3}{5}$ Ni and $\frac{3}{5}$ Sr. The factors, f_c and f_w were estimated from standard beta absorption curves and are accurate to $\pm 2\%$. The factor, f_{b} , was determined for aluminum and stainless steel planchets and it is accurate to $\pm 1\%$. The values for f_{ssa} were derived from the dat of Nervik and Stevenson (7) and are accurate to $\pm 5\%$. The product of all the corrections factors is accurate to within $\pm 7\%$.

58Co had to be treated somewhat differently since 85% of its decays are by electron capture and 15% are by positron emission (8). To obtain a self-absorption correction factor for the x-rays, (I/I_0) , the following expression was used:

$$
\left(\frac{I}{I_0}\right) = \frac{\omega_K \left(\frac{K}{K+L}\right)}{(\rho x) (\mu/\rho)} \left[1 - e^{-(\mu/\rho)(\rho x)}\right]
$$

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where ρx = the CoS sample thickness in mg/cm²;

 (μ/ρ) = mass attenuation coefficient = 0.106 cm²/mg ⁽²⁾; (μ / ρ) - mass attenuation coefficient = 0.106 cm /mg μ ;
 $\left(\frac{K}{\sigma}\right)$ = ratio of the number of K x-rays to the total number $K +$ of x-rays emitted = 0.90 (8)

 $\omega_{\rm K}$ = fluorescence yield for Fe x-rays = 0.293 for K; 0.0 for L (8) Auger electrons will not leave the sample or penetrate the cover since their range is $\angle 0.2$ mg/cm² . Thus the "feator for $58\degree$ can be calssa culated by ${}^{\text{II}}\mathbf{f}_{\text{SSa}} = 0.15(f_{\text{SSa}})_{\text{R}} + + 0.85(I/\mathbf{I}_{\Omega}^{\bullet}).$ This \mathbb{I} I' value is then used in equation (i) to calculate the total ssa activity of 58 Co.

To obtain the corrections factors for the gross beta samples, it is standard to treat them as though they were 137 Cs samples (4) even though they are a mixture of many activities.

The detection limits for this system, assuming a 50% yield of each fraction, is summarized in Table 2.

EXPERIMENTAL RESULTS

The concentrations of the various nuclides in Dardenelle Lake water samples are listed in Tables 3 - 10.

(2) This value is obtained by averaging the (μ/ρ) values of Co (0.80) and S (0.173) (9)

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Tables $3 - 10$: Concentrations

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Selected Nuclides

- NOTES: (a) Sample collected at Station 5
	- (b) Sample collected at Station 16
	- (c) Sample collected at Station 21
	- (d) Sample collected at Station 7
	- (e) Sample of very low chemical yield

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Table 3: 144 Ce $^{-144}$ Pr Concentrations (pCi/1 x 10²)

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Table 5: 58 Co Concentrations (pCi/1 x 10²)

Table 7: 90 Sr - 90 Y Concentrations (pCi/l x 10²)

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Table 9: 110m Ag Concentration (pCi/1 x 10²)

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Table 10: Gross Beta Concentration (pCi/l)

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CONCLUSIONS

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- 1. The 144 Ce 144 Pr concentrations seem little affected by reactor operation, although concentrations increased at every station sampled after the reactor had commenced full operation after refueling in the late winter of 1977. There are two possible explanations for this increase: (a) a natural increase at all stations due to fallout; (b) since the flow rate of the water through the dam and hydroelectric power station was generally quite low during most of the project period, release from Nuclear I followed by good mixing in the lake, could cause this general increase. Even so, the concentrations of the nuclides were consistently very low.
- 2. The 141 Ce and 90 Sr found in the ten samples were formed predominantly by Chinese nuclear testing in the atmosphere. The first of the three tests in the autumn of 1976 seems to be the dirtiest. There seems to be two significant injections of these nuclides into the lake water, one in the autumn of 1976 and the other in April-May, 1977 (see 6 below).
- 3. Amounts of 58 Co in excess of 0.12 pCi/l appear only at station 6 between December, 1975 and August, 1976 and at most stations in the summer of 1976. There is no natural explanation for this very systematic long term appearance of relatively large amounts of 58 Co at station 6. It is interesting to note that the concentration decreases with a half-life of 72 days. In some way, this sampling area was contaminated with 58 Co and the contamination did not move with time.
- 4. In June, July and August, 1976, as the result of a planned release, unusually high concentrations of 58 Co were found in many samples. It will be noted that the concentration decreases slowly with time. This is

a result of the slow turnover of water during the dry spell of 1976 and 1977.

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It was also noted that the dilution between stations 1 and 4 (some 2-3 miles distant) was only \sim 3-fold. This indicates very little mixing on the sampling day, but, over the long range, there seems to be significant mixing in the reservoir as a whole. This is borne out by the 137 Cs data. 5. The 137 Cs concentrations are somewhat affected by day-to-day reactor operation but the only significant injections of 137_{Cs} were in June, 1976 and in January, 1977. The trends mentioned in 3, above, can be seen also with 137 Cs. This dilution on release day was only \sim 14-fold and the effects of the release were seen for several months.

6. The 90 Sr- 90 Y concentrations remained fairly constant, with a few exceptions, until January, 1977. At that time Nuclear I was shut down for refueling. During the down period, the majority of samples showed a marked decrease in 90 Sr- 90 activity. Refueling was completed in late March and the activity in the following months showed a dramatic and regular increase. Unusually high activities began to sppear erratically at Stations 2 and 3 (the upstream stations). The 89 Sr concentrations also followed this erratic pattern, so it was tempting to ascribe the high concentrations in the spring of 1977 to fallout. Analysis of the increases in 89 Sr and 90 Sr, though, showed that this cannot be the case. 89 Sr began to appear with the fallout from the Chinese tests but there was very little change in the 90_{ST} , since $\rm A(^{90}Sr)$ \leq $\rm A(^{89}Sr)$ in the newly produced fallout even a year after the tes ts .

If the source of contamination is cooling water, the low ${}^{89}Sr/{}^{90}Sr$ ratio indicated that the contaminated water was stored for some time before release.

The Sr contamination could have also come from the outgassing of volatile precursors (89) Kr and 90 Kr) into the atmosphere. This could account for the widespread appearance of increased activities, both upstream and downstream.

These increases in Sr activity coincide with increases in 137 Cs in many samples during the spring of 1977. ¹³⁷Cs also has a volatile precursor, ¹³⁷Xe.

Reactor operation is causing an increase in the 90 Sr- 90 Y activity in Lake Dardenelle by a factor of 1.5-3. Reactor operation may also have been the cause of the increase of 89 Sr activity in the spring of 1977.

- There is no discernible pattern to the 110m Ag concentration except 7. to note its appearance in appreciable amounts in the planned release of June 21, 1976 and in a few samples in the winter of 1975-1976.
- It was a good deal more difficult to analyze the gross β data. 8. Winter samples usually contain smaller amounts of solids and thus lower activities. Sample sizes were also more variable than in summer. Graphical analysis lends itself to these data. Figure 2 is a plot of the gross β activity (pCi/1) vs. the mass of unfiltrable solids per liter of water. It can be seen that many points fall on or near the solid line. The dashed lines are the $\pm 10\%$ limits which

Figure 2: Total Gross 8 Activity Per Liter of Water as a Function of the Mass of the Unfi1trab1e Solids

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 $x - upstream$ stations $(2, 36, 16, 21)$

 \blacklozenge - downstream stations $(1, 4, 5, 7)$

Two points for the samples of 6/21/76 have been omitted because they are off the activity scale on graph.

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Gross 8 Activity (pCi/I)

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approximate the standard deviation. The equation for this line can be expressed as:

 $A = (sp. act.)m + b$

where $A =$ activity of the gross β sample (pCi/1)

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 $(sp. act.) = the specific activity of naturally occurring$ radioactivity (pCi/mg)

 $m =$ mass of unfiltrable solids per liter of water (mg/1)

 $b =$ activity of other radio active nuclides that did not result from reactor operation (e.g. fallout)

Any point above the solid line represents a sample which, in addition to natural radioactivity (mostly 40 K) and radioactivity from fallout. contains some radioactive nuclides that are reactor produced.

If this approach is valid, the slope (sp. act.) should agree with the specific activity of 40 K in river water and the intercept (b) should correspond to the total activity of the fallout. The value for the slope was determined to be $\approx 1.0 \times 10^{-2}$ pCi/mg and the intercept to be $\sim 0.7 - 0.9$ pCi/1.

If it is assumed that all the naturally occurring activity is due to the decay of 40 K, the unfiltrable solids would contain 1.1% K by weight.

Mason (10) cites a value of $\sim 2\%$ for the K content of dissolved solids in river water, twice the percentage found in Dardenelle Lake water. The low value may be explained by the fact that the Ca content of the Arkansas River is much higher than the value for Ca quoted by Mason The intercept is also of the right magnitude. 90_{ST} - 90_{Y} is the major radionuclide in samples unaffected by reactor effluent and its concentration is normally in the range $0.7 - 0.9$ pCi/1.

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It is safe to assume that points above the A vs. mass line indicate the presence of reactor generated nuclides. An appreciable number of samples, from both upstream and downstream stations show this excess activity.

It will be noticed that some points lie significantly below the line. The percentage of K must be lower here than in the average sample. These points are all for upstream samples, most of them taken from station 2, where the water is moving quite swiftly and perhaps carrying a larger colloidal load than at other sampling sites.

- 9. Overall, it can be concluded that the day-to-day operation of Nuclear I has very little effect on the radionuclide concentration in Dardenelle Lake water. The fallout from the Chinese tests was more evident. Only planned releases caused significantly increased concentrations. Dilution by river water may not be as effective as could be wished but this may be a seasonal effect. The slow turnover of lake water during dry periods also poses a problem because of the spread of reactor produced radionuclides throughout the lake.
- 10. It must be pointed out that, at no time, did the concentrations of the measured nuclides exceed the limits set by the Arkansas State Department

of Health. Concentrations were so low that it was decided not to pursue a study of the uptake of radionuclides in biological systems or in their deposition with sediments.

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